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(54) MAGNETIC POLYMER NANOCOMPOSITE FOR REMOVAL OF DIVALENT HEAVY METAL IONS FROM WATER

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(58) Field of Classification Search CPC B01J 20/261 See application file for complete search history.

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(57)ABSTRACT

The magnetic polymer nanocomposite for removal of divalent heavy metal ions from water is magnetic nanocomposite having a core of magnetite (Fe₃O₄) in a shell of branched polyhydroxystyrene (BHPS), designated as Fe₃O₄@BHPS. The nanocomposite is synthesized by co-precipitation in alkali solution. Testing showed the nanocomposite reached 93% and 80% Pb(II) and Cd(II) adsorption, respectively, in 30 minutes, attaining equilibrium in 120 minutes. The maximum adsorption capacities of Pb(II) and Cd(II) at 298K were 186.2 and 125 mg/g, respectively. After adsorption, the nanocomposite with the heavy metal(s) adsorbed thereto was easily removed from aqueous solution by application of a magnetic field.

5 Claims, 9 Drawing Sheets



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FIG. 4



FIG. 5A



FIG. 5B











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MAGNETIC POLYMER NANOCOMPOSITE FOR REMOVAL OF DIVALENT HEAVY METAL IONS FROM WATER

BACKGROUND

1. Field

The disclosure of the present patent application relates to magnetic nanocomposite material, and particularly to a ¹⁰ magnetic polymer nanocomposite for removal of divalent heavy metal ions, such as lead and cadmium, from water.

2. Description of the Related Art

Rapid industrialization and population growth drive an increasing need for, and threaten the fate of, clean water availability. Environmental protocols in many developed countries have changed to sustain the environment and conserve water reservoir quality. Among potential water ²⁰ contaminants, heavy metals are of foremost concern, as they are persistent, potentially toxic and carcinogenic, even in trace amounts. Therefore, monitoring and controlling heavy metal concentrations in waste streams before discharge to surface and subsurface water is essential. ²⁵

Diverse water treatment technologies (for example, chemical precipitation, electro-deposition, ultra-filtration, ion-exchange, adsorption and biological processes) have been engineered to remove heavy metals present in water. Operational ease, cost-effectiveness and applicability, even 30 at very low adsorbate concentrations, make adsorption a highly acclaimed water remediation technique. Advancements in nanotechnology have opened many new avenues in the development of nanosized adsorbents. The large surface area to volume ratio and high adsorption efficiency of 35 nanoadsorbents make them strong competitors to conventional adsorbents for use in water purification. However, while nanoadsorbents are potentially effective water remediation agents, their small particle size presents practical obstacles, such as difficulty separating saturated nanoadsor- 40 bents from unsaturated nanoadsorbents or the aquatic milieu being treated, restricting their utility in water treatment applications.

Magnetic separation processes were introduced in the early 1990's to address such phase separation issues, and ⁴⁵ have since gained considerable attention. During magnetization separation, nanoadsorbent particles are embedded, encapsulated or coated with metal oxides (for example, magnetite, Fe_3O_4) to introduce magnetic properties.

Unexplored in the context of magnetic separation pro- ⁵⁰ cesses and related materials, branched polyhydroxystyrene (BPHS), a derivative of hydroxystyrene, is widely used in HPLC columns as an organic waste adsorbent, and in selectively permeable membranes.

Thus, a magnetic polymer nanocomposite for the removal ⁵⁵ of divalent heavy metal ions from water solving the aforementioned problems is desired.

SUMMARY

The magnetic polymer nanocomposite for removal of divalent heavy metal ions from water is a magnetic nanocomposite having a core of magnetite (Fe₃O₄) in a shell of branched polyhydroxystyrene (BHPS), designated as Fe₃O₄@BHPS. The nanocomposite is synthesized by co- 65 precipitation in alkali solution. Testing showed the nanocomposite reached 93% and 80% Pb(II) and Cd(II) adsorp-

tion, respectively, in 30 minutes, attaining equilibrium in 120 minutes. The maximum adsorption capacities of Pb(II) and Cd(II) at 298K were 186.2 and 125 mg/g, respectively. After adsorption, the nanocomposite with the heavy metal(s) adsorbed thereto was easily removed from aqueous solution by application of a magnetic field.

A method of removing a divalent heavy metal ions from aqueous solutions includes the steps of: (a) adding a magnetic polymer nanocomposite having a magnetite core and a branched polyhydroxystyrene shell (Fe3O4@BPHS) to an aqueous solution contaminated with a divalent heavy metal ions; (b) leaving the magnetic polymer nanocomposite in the aqueous solution for a period of time sufficient to saturate the magnetic polymer nanocomposite with the divalent heavy metal ions adsorbed on the nanocomposite; and (c) removing the saturated magnetic polymer nanocomposite from the aqueous solution by applying an external magnetic field. The method may further include the steps of: (d) eluting the magnetic polymer nanocomposite having the divalent heavy metal ions adsorbed thereon with a dilute solution of a strong acid in order to desorb the divalent heavy metal ions from the magnetic polymer nanocomposite; and (e) applying an external magnetic field to recover the magnetic polymer nanocomposite free from adsorbed divalent heavy metal ions for re-use.

These and other features of the present disclosure will become readily apparent upon further review of the following specification and drawings.

BRIEF DESCRIPTION OF THE DRAWINGS

FIG. **1** is a reaction scheme for the synthesis of a magnetic polymer nanocomposite for removal of divalent heavy metal ions from water.

FIG. **2** is a reaction scheme diagrammatically showing the adsorption of lead [Pb(II)] and cadmium [Cd(II)] ions by the magnetic polymer nanocomposite synthesized in FIG. **1**.

FIG. 3 is a composite of the Fourier transform infrared spectroscopy (FTIR) spectra of the intermediate products, viz., magnetite (Fe₃O₄) and branched polyhydroxystyrene (BPHS), and the product magnetic polymer nanocomposite of the reaction shown in FIG. 1.

FIG. 4 is a plot of the X-ray diffraction (XRD) patterns of Fe_3O_4 and the magnetic polymer nanocomposite, $Fe_3O_4@BHPS$.

FIG. **5**A is a scanning electron micrograph (SEM) of branched polyhydroxystyrene (BPHS), and FIG. **5**B is a SEM micrograph of the magnetic polymer nanocomposite.

FIG. **6**A is an energy dispersive X-ray (EDX) plot of the magnetic polymer nanocomposite; FIG. **6**B is an energy dispersive X-ray (EDX) plot magnetic polymer nanocomposite after Pb(II) adsorption; and FIG. **6**C is an energy dispersive X-ray (EDX) plot of the magnetic polymer nanocomposite after Cd(II) adsorption.

FIG. **7** is a thermogravimetric analysis (TGA)/differential thermal analysis (DTA) plot of the magnetic polymer nanocomposite.

FIG. **8**A is a composite plot showing adsorption capacity as a function of initial pH for the adsorption of Pb(II) and Cd(II) by the magnetic polymer nanocomposite.

FIG. **8**B is a composite plot showing adsorption capacity as a function of contact time for the adsorption of Pb(II) and Cd(II) by the magnetic polymer nanocomposite.

FIG. **8**C is a composite plot showing adsorption capacity as a function of equilibrium concentration for the adsorption of Pb(II) by the magnetic polymer nanocomposite at different temperatures.

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FIG. 8D is a composite plot showing adsorption capacity as a function of equilibrium concentration for the adsorption of Cd(II) by the magnetic polymer nanocomposite at different temperatures.

FIG. 9A is a chart showing regeneration of the magnetic polymer nanocomposite saturated with Pb(II) heavy metal ions by elution with 0.01M solutions of hydrochloric acid (HCl), nitric acid (HNO₃), and sulfuric acid (H₂SO₄), respectively.

FIG. 9B is a chart showing regeneration of the magnetic 10polymer nanocomposite saturated with Cd(II) heavy metal ions by elution with 0.01M solutions of hydrochloric acid (HCl), nitric acid (HNO₃), and sulfuric acid (H_2SO_4), respectively.

Similar reference characters denote corresponding fea- 15 tures consistently throughout the attached drawings.

DETAILED DESCRIPTION OF THE PREFERRED EMBODIMENTS

The magnetic polymer nanocomposite for removal of divalent heavy metal ions from water is magnetic nanocomposite having a core of magnetite (Fe₃O₄) in a shell of branched polyhydroxystyrene (BHPS), designated as Fe₃O₄@BHPS. The nanocomposite is synthesized by co- 25 precipitation in alkali solution. Testing showed the nanocomposite reached 93% and 80% Pb(II) and Cd(II) adsorption, respectively, in 30 minutes, attaining equilibrium in 120 minutes. The maximum adsorption capacities of Pb(II) and Cd(II) at 298K were 186.2 and 125 mg/g, respectively. 30 After adsorption, the nanocomposite with the heavy metal(s) adsorbed thereto was easily removed from aqueous solution by application of a magnetic field.

A method of removing a divalent heavy metal ions from aqueous solutions includes the steps of: (a) adding a mag- 35 netic polymer nanocomposite having a magnetite core and a branched polyhydroxystyrene shell (Fe₃O₄@BPHS) to an aqueous solution contaminated with a divalent heavy metal ions; (b) leaving the magnetic polymer nanocomposite in the aqueous solution for a period of time sufficient to saturate 40 the magnetic polymer nanocomposite with the divalent heavy metal ions adsorbed on the nanocomposite; and (c) removing the saturated magnetic polymer nanocomposite from the aqueous solution by applying an external magnetic field. The method may further include the steps of: (d) 45 eluting the magnetic polymer nanocomposite having the divalent heavy metal ions adsorbed thereon with a dilute solution of a strong acid in order to desorb the divalent heavy metal ions from the magnetic polymer nanocomposite; and (e) applying an external magnetic field to recover the 50 magnetic polymer nanocomposite free from adsorbed divalent heavy metal ions for re-use.

The magnetic polymer nanocomposite for removal of divalent heavy metal ions from water will be better understood with reference to the following examples.

Example 1

Synthesis of Magnetic Polymer Nanocomposite (Fe₃O₄@BPHS)

In an exemplary synthesis, 2.36 g ferric chloride hexahydrate (FeCl₃.6H₂O) was dissolved in 100 mL 0.5M HCl solution in a 250 mL Erlenmeyer flask. In another 250 mL Erlenmeyer flask, 0.86 g ferrous chloride (FeCl₂.4H₂O) was 65 dissolved in 100 mL 0.5M HCl solution under N₂ gas flow for 20 min to prevent the oxidation of Fe(II). Then, the two

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solutions were mixed in a three-necked round bottom flask equipped with a mechanical stirrer at 80° C. under N2 gas flow to form a reaction mixture. After mixing, 0.5 g of branched polyhydroxystyrene (BPHS), in powder form, was dissolved in 30 mL methanol (CH₃OH), and the methanolic BPHS solution was added dropwise, together with 20 mL of 28% NH₂OH solution, to the reaction mixture under rigorous mechanical stirring (1000 rpm) for a period of 30 min until the pH of the reaction mixture reached around 10. The reaction mixture was then stirred for another 4 hrs under N₂ atmosphere. Finally, a resulting black colloidal product was separated from the reaction mixture by application of an external magnetic field, and washed with deionized water and methanol to remove unreacted BPHS and other impurities. The precipitate was dried at room temperature over a watch glass for 48 hrs to obtain BPHS impregnated with magnetite nanoparticles, i.e., Fe₃O₄@BPHS. The synthesis scheme is shown diagrammatically in FIG. 1.

The resulting product was characterized by Fourier transform infrared spectroscopy (FTIR), the spectra of the reactants and the products being compared in FIG. 3. In the spectrum of BPHS, a broad absorption peak is observed at around 3407 cm⁻¹ (presumably stretching of -OH); 3022 cm^{-1} (presumably stretching of Ar—C—H); 2963 cm⁻¹ and 2831 cm⁻¹ (presumably asymmetric and symmetric stretching of $-\tilde{C}$ —H, respectively); 1604 cm⁻¹ (stretching) Ar-C=C); 1507, 1448, and 1237 cm⁻¹ (bending -OH, and -C-O). In the spectrum of Fe₃O₄@BPHS nanocomposite, 572 cm⁻¹ (Fe—O); 3398 cm⁻¹ (stretching —OH); 3033 (stretching Ar—C—H); 2936 cm⁻¹ and 2838 (asymmetric and symmetric stretching —C—H, respectively); 1603 (stretching Ar-C=C), 1514, 1398, and 1220 cm⁻ (bending —OH, and —C—O).

The resulting product was also characterized by X-ray diffraction (XRD) the magnetite and the final product patterns being compared in FIG. 4. Observed characteristic peaks at 20=18.33°, 30.2°, 35.3°, 43.2°, 53.5°, 57.3°, 63.2° and 74.9° correspond to (111), (220), (311), (400), (422), (511), (440) and (533) indices, respectively, consistent with standard results for Fe₃O₄ in the Joint Committee on Powder Diffraction Standards database (JCPDS, file 19-0629), confirming the cubic spinel structure of pure Fe₃O₄. XRD results for Fe₃O₄@BPHS were qualitatively similar to those of Fe₃O₄, with no new peaks observed.

The crystal sizes of Fe₃O₄ and Fe₃O₄@BPHS were determined from their respective XRD patterns by using Debye-Scherrer's equation, provided as equation (1):

$$D_{S} = K \times \frac{\lambda}{(\beta \cos \theta)},\tag{1}$$

where θ is the Bragg's peak, β is the half-width of the 55 diffraction peak, K is a constant that depends on the crystallite's morphology, and λ is the X-ray wavelength, assumed to be 0.154 nm. The (311) peak of the highest intensity was picked out to evaluate the crystal size. The crystallite sizes of Fe₃O₄ and Fe₃O₄@BPHS were found to be 11.8 and 14.5 nm, respectively.

In addition, the resulting product was also characterized by SEM microscopy, FIGS. 5A and 5B comparing the BPHS reactant with the Fe₃O₄@BPHS product. A highly porous and uneven BPHS surface with no uniform sized pores was observed, as shown in FIG. 5A. FIG. 5B shows the Fe₃O₄@BPHS appears to take the form of a BPHS sheet impregnated with magnetite nanoparticles.

Finally, the resulting product was characterized by thermogravimetric analysis (TGA)/differential thermal analysis (DTA), a plot of the product being shown in FIG. **7**. A two-step weight loss within the temperature range 30-800° C. was observed. The first-step shows 6% weight loss due to the loss of physically absorbed water molecules, observed in the temperature range 30-280° C. The second-step shows 37% weight loss due to decomposition of the BPHS polymer component of the Fe₃O₄@BHPS, and was observed within a temperature range of 325-550° C.

The surface area, pore volume and pore size of the exemplary Fe_3O_4 @BHPS were determined through adsorption of nitrogen (N₂) gas at 77 K using a Micromeritics Surface Area and Porosity analyzer (Gemini VII, 2390 Surface Area and Porosity analyzer). Prior to BET surface area measurement, the sample was degassed at 150° C. for 2 h. The surface area analysis of the exemplary Fe_3O_4 @BPHS showed a Brunauer-Emmett-Teller (BET) surface area of 41.8 m²/g and an average pore diameter of 1.9 nm.

Example 2

Adsorption Experiments and Results

Batch mode lead [Pb(II)] and cadmium [Cd(II)] adsorption experiments were carried out to test the efficacy of the Fe₃O₄@BPHS, synthesized as described in Example 1, as an adsorbent. In a 100 mL Erlenmeyer flask, 25 mL Pb(II) aqueous solution having an initial concentration (C_o) of 25 ³⁰ mg/L was equilibrated with 0.015 g Fe₃O₄@BPHS over a thermostatic shaking assembly at 100 rpm operated at room temperature. At equilibrium, the saturated Fe₃O₄@BPHS nanocomposite was separated from the aqueous phase by an external magnetic field and the residual Pb(II) concentration ³⁵ was quantitatively determined by atomic adsorption spectrophotometry (AAS). Experimental parameters of pH, contact time, C_o and temperature were varied. Similar batch mode adsorption experiments were conducted for Cd(II).

The amount $(q_e, mg/g)$ and percentage (%) of Pb(II) and ⁴⁰ Cd(II) adsorption were calculated according to equation (2) and equation (3), respectively, as follows:

$$q_{e}, \operatorname{mg/g} = (C_{o} - C_{e}) \times \frac{V}{m}$$
⁽²⁾ 45

% adsorption =
$$\frac{C_o - C_e}{C_o} \times 100,$$
 (3)

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where V is the volume of adsorbate solution (in L), m is the mass of $\text{Fe}_3\text{O}_4(@BPHS \text{ added (in g)}, \text{ and } \text{C}_o \text{ and } \text{C}_e$ (in mg/L) are the initial and equilibrium concentrations of Pb(II) or Cd(II) ions in solution, respectively.

FIG. **2** is a reaction scheme diagrammatically showing the adsorption of lead [Pb(II)] and cadmium [Cd(II)] ions by the magnetic polymer nanocomposite synthesized in FIG. **1**. FIG. **6**A is an energy dispersive X-ray (EDX) plot of the magnetic polymer nanocomposite; FIG. **6**B is an energy dispersive X-ray (EDX) plot magnetic polymer nanocomposite after Pb(II) adsorption; and FIG. **6**C is an energy dispersive X-ray (EDX) plot of the magnetic polymer nanocomposite after Cd(II) adsorption. FIG. **6**A shows C, O and Fe present in pristine Fe₃O₄@BPHS. FIGS. **6**B and **6**C show the presence of Pb(II) and Cd(II) in the Fe₃O₄@BPHS following adsorption experiments, respectively, confirming binding of the adsorbed heavy metals to Fe₃O₄@BPHS surface.

FIG. 8A is a composite plot showing adsorption capacity as a function of initial pH for the adsorption of Pb(II) and Cd(II) by the magnetic polymer nanocomposite. Maximal Pb(II) and Cd(II) ion adsorption on the Fe₃O₄@BPHS nanocomposite was achieved for initial pH (pH_i) values above 4 and 6, respectively; preferably between 5-7 and 6.5-9, respectively; and more preferably around 5.5 and 7.1, respectively. FIG. 8B is a composite plot showing adsorption capacity as a function of contact time for the adsorption of Pb(II) and Cd(II) by the magnetic polymer nanocomposite. The equilibration times for Pb(II) and Cd(II) ions adsorption on the Fe₃O₄@BPHS, i.e., the times for saturation, were at least 30 and 90 minutes, respectively; and preferably at least 90 and 120 min, respectively. FIG. 8C is a composite plot showing adsorption capacity as a function of equilibrium concentration for the adsorption of Pb(II) by the magnetic polymer nanocomposite at different temperatures. FIG. 8D is a composite plot showing adsorption capacity as a function of equilibrium concentration for the adsorption of Cd(II) by the magnetic polymer nanocomposite at different temperatures. The adsorption of Pb(II) and Cd(II) ions onto Fe₃O₄@BPHS was found to slightly decrease with increasing temperature over the range from 298-318K. The maximum capacity for Pb(II) and Cd(II) ion adsorption onto Fe₃O₄@BPHS was measured in the present exemplary experiments to be 186.2 mg/g and 125 mg/g, respectively at 298 K.

Table 1 is the isotherm data for the adsorption of Pb(II) and Cd(II) on the magnetic polymer nanocomposite, $Fe_3O_4@BPHS$ synthesized in Example 1.

Table 2 is the kinetic data for the adsorption of Pb(II) and Cd(II) on the magnetic polymer nanocomposite, Fe3O4@BPHS when $C_o=25 \text{ mg/L}$.

TABLE 1	
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Isotherm data for adsorption of Pb(II) and Cd(II)							
	Isotherm models						
	Langmuir Freundlich						
Temperature (K)	q _m (mg/g)	b (L/mg)	R _L	R ²	$\frac{\mathrm{K}_{F}}{(\mathrm{mg/g})(\mathrm{L/mg})^{1/n}}$	n	R ²
298 308	163.9	0.085	0.319	0.964	19.64	2.10	0.981
308 318 298 308 318	138.7 129.9 101.0 72.5 63.3	0.038 0.064 0.121 0.147 0.153	0.465 0.438 0.292 0.254 0.246	0.962 0.904 0.976 0.888 0.834	14.07 12.49 20.39 17.29	2.00 1.94 2.92 3.21 3.34	0.973 0.972 0.989 0.989
	Temperature (K) 298 308 318 298 308 318	Isotherm Temperature (K) qm (mg/g) 298 163.9 308 158.7 318 129.9 298 101.0 308 72.5 318 63.3	Isotherm data for a Lang Temperature (K) q _m b 298 163.9 0.085 308 158.7 0.058 318 129.9 0.064 298 101.0 0.121 308 72.5 0.147 318 63.3 0.153	Isotherm data for adsorption I	$\begin{tabular}{ c c c c c c c c c c c c c c c c c c c$	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{tabular}{ c c c c c c } \hline Isotherm data for adsorption of Pb(II) and Cd(II) \\ \hline Isotherm models \\ \hline Isotherm models \\ \hline Isotherm models \\ \hline Langmuir & Freundlich \\ \hline Langmuir & K_F \\ \hline (mg/g) & (L/mg) & R_L & R^2 & (mg/g)(L/mg)^{1/n} & n \\ \hline 298 & 163.9 & 0.085 & 0.319 & 0.964 & 19.64 & 2.10 \\ \hline 308 & 158.7 & 0.058 & 0.465 & 0.962 & 14.67 & 2.00 \\ \hline 318 & 129.9 & 0.064 & 0.438 & 0.904 & 12.49 & 1.94 \\ \hline 298 & 101.0 & 0.121 & 0.292 & 0.976 & 20.39 & 2.92 \\ \hline 308 & 72.5 & 0.147 & 0.254 & 0.888 & 17.29 & 3.21 \\ \hline 318 & 63.3 & 0.153 & 0.246 & 0.834 & 15.94 & 3.34 \\ \hline \end{tabular}$

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Kinetic data for adsorption of Pb(II) and Cd(II)							
	Kinetic models						
		Ps	Pseudo-first-order Pseudo-second-order				
Metal ions	q _{e,exp.} (mg/g)	q _{e,cal.} (mg/g)	k ₁ (1/min)	R ²	q _{e,cal.} (mg/g)	k ₂ (g/mg-min)	R ²
Pb(II) Cd(II)	42.3 41.2	163.9 101.0	0.085 0.121	0.937 0.929	20.39	2.10 2.92	0.981 0.989

Example 3

Regeneration Experiments and Results

Pb(II) and Cd(II) ions were desorbed from saturated Fe_3O_4 @BPHS nanocomposite by elution by a strong acid, such as HCl, H₂SO₄, or HNO₃, in batch mode experiments. 20 ite, comprising the steps of: Samples of Fe₃O₄@BPHS, saturated with a divalent heavy metal ion $(Pb^{+2} \text{ or } Cd^{+2})$ removed from water or aqueous solution as described in Example 2, was treated with a 25 mL 0.01 M solution of one of the various eluents over a thermostatic shaker assembly shaken at 100 rpm for 24 hr to 25 ensure complete elution of the adsorbed heavy metal ions by the acid. The solid and solution phases of the resulting mixture were magnetically separated, and the eluted heavy metal ions concentration was quantitatively determined by atomic absorption spectroscopy (AAS). The amount of 30 heavy metal ions desorbed was calculated according to equation (3):

% Desorption =
$$\frac{C_d}{C_{ia}} \times 100,$$
 (4)

where C_d is the concentration of the heavy metal ions desorbed by the eluent and C_{ia} is the initial concentration of 40 heavy metal ions adsorbed on the magnetic polymer composite, Fe₃O₄@BPHS. FIGS. 9A and 9B show the results of eluting the used magnetic polymer nanocomposite with a dilute solution of a strong acid.

An adsorption/desorption scheme of Pb(II) and Cd(II) on Fe₃O₄@BPHS is shown in FIG. 2 using hydrochloric acid as an example, although 0.1 M nitric acid or 0.1 M sulfuric acid may be used instead of hydrochloric acid, as shown by FIGS. 9A and 9B.

It is to be understood that the magnetic polymer nanocomposite for removal of divalent heavy metal ions from water is not limited to the specific embodiments described above, but encompasses any and all embodiments within the scope of the generic language of the following claims enabled by the embodiments described herein, or otherwise shown in the drawings or described above in terms sufficient to enable one of ordinary skill in the art to make and use the claimed subject matter.

We claim:

1. A method of making a magnetic polymer nanocompos-

- dissolving FeCl₂ and FeCl₂ separately in HCl to form an Fe(III) solution and an Fe(II) solution, respectively;
- combining the Fe(III) solution and the Fe(II) solution to form a combined solution;
- adding a solution of branched polyhydroxystyrene (BPHS) in methanol and a solution of ammonium hydroxide to the combined solution dropwise to form a mixture; and
- stirring the mixture under nitrogen atmosphere to precipitate the magnetic polymer nanocomposite as Fe₃O₄ nanoparticles in a BPHS shell (Fe₃O₄@ BPHS).

2. The method of making a magnetic polymer nanocomposite according to claim 1, further comprising the step of removing the precipitated magnetic polymer nanocomposite ³⁵ by application of an external magnetic field.

3. The method of making a magnetic polymer nanocomposite according to claim 1, wherein said step of adding a solution of branched polyhydroxystyrene (BPHS) in methanol and a solution of ammonium hydroxide to the combined solution dropwise comprises adding a solution of branched polyhydroxystyrene (BPHS) in methanol and a solution of ammonium hydroxide to the combined solution dropwise until a pH of 10 is obtained.

4. The method of making a magnetic polymer nanocomposite according to claim 1, wherein said step of dissolving FeCl₃ and FeCl₂ separately in HCl to form an Fe(III) solution and an Fe(II) solution further comprises dissolving the FeCl₂ in HCl in the presence of a flow of nitrogen gas to prevent oxidation of Fe(II).

5. The method of making a magnetic polymer nanocomposite according to claim 1, wherein said solution of ammonium hydroxide comprises a 28% NH₄OH solution.

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